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大型真菌对镉的耐受机制研究进展

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摘要 镉(Cd)是毒性最强的重金属之一,目前也是我国土壤的首要污染物,Cd可以在土壤和生物体中积累,影响生态系统,并通过食物链威胁人类健康,土壤Cd污染的修复已成为迫切需要解决的问题.总结土壤Cd污染现状和大型真菌对Cd的富集作用;重点从生理生化和转录水平上梳理大型真菌对Cd的耐受机制,主要包括:胞外配体的结合降低其活动性和细胞壁组分的结合共同减少真菌细胞对Cd的吸收,胞内抗氧化系统响应以减少Cd胁迫带来的氧化损伤,Cd-螯合物形成及区室化作用隔离Cd以降低其毒性;细胞壁修饰、抗氧化物质合成、跨膜转运蛋白和金属结合相关基因上调可能是大型真菌对Cd富集和耐受的重要途径;这些可为大型真菌耐Cd机制的系统阐明提供重要参考.大型真菌对Cd具有极高的富集能力及耐受性,作为修复材料在土壤Cd污染修复方面具有应用价值.但大型真菌对于Cd的耐性机制存在种和菌株的特异性,单一的机制并不能解释大型真菌对Cd的耐受性,要将真菌修复应用于Cd污染土壤修复实践,未来需要强化生理生化试验与多组学技术的联合应用以系统地阐明大型真菌对Cd胁迫的富集过程和耐受机制存在的共性及特性;加强对具有富集Cd能力的大型真菌的筛选、驯化及人工种植研究,丰富修复材料的可选择性;探索废弃物的后续利用,实现资源的循环利用. (图1 表1 参112)

关键词 土壤; Cd污染; 真菌修复; 耐受机制; 转录组学

Tolerance mechanism of macrofungi to cadmium: a review

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Abstract Cadmium (Cd) is one of the most toxic heavy metals and is currently the primary soil pollutant in China. Cd can accumulate in soil and organisms, harming ecosystems and threatening human health throughout the food chain. The remediation of Cd pollution in soil has become an urgent problem that needs to be solved. This review summarizes the current status of soil Cd pollution and the enrichment of Cd by macrofungi. The focus of this review was to summarize the mechanisms of Cd tolerance in macrofungi at the physiological, biochemical, and transcriptional levels, which mainly include the combination of extracellular ligands to reduce Cd activity and the combination of cell wall components to jointly reduce the absorption of Cd by fungal cells, the response of the intracellular antioxidant system to reduce the oxidative damage caused by Cd stress, and the formation of Cd-chelate and compartmentalization to isolate Cd to reduce its toxicity. Upregulation of cell wall modification, antioxidant synthesis, transmembrane transport proteins, and metal-binding-related genes may be important pathways for the accumulation and tolerance of Cd in macrofungi. This is important for providing a reference for the systematic elucidation of Cd tolerance mechanisms in macrofungi. These organisms have an extremely high accumulation ability and tolerance for Cd and have applications in the remediation of soil Cd pollution. However, there are species and strain specificities in the Cd tolerance mechanism of macrofungi, and a single mechanism cannot explain their tolerance to Cd. To apply fungal remediation to Cd-contaminated soil, it is necessary to strengthen the combined application of physiological and biochemical experiments and multi-omics technology in the future to systematically clarify the commonality and characteristics of the enrichment process and tolerance mechanism of macrofungi to Cd, strengthen the screening, domestication, and artificial cultivation of macrofungi with the ability to enrich Cd for the rich selection of repair materials, and explore the follow-up utilization of waste to realize the recycling of resources.

Keywords soil; Cd pollution; fungal remediation; tolerance mechanism; transcriptomics

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镉(Cd)是毒性最强和农田受污染最普遍的重金属之一,Cd对植物种子萌发、生长、产量和生物量具有明显的毒害作用^[1-3],且可通过食物链进入人体,在人体中蓄积,对肾脏、骨骼和肺等器官影响巨大^[4],并具有致癌效应^[5,6]。2014年4月公布的《全国土壤污染状况调查公报》显示,我国耕地、林地、草地、未利用土地土壤点位超标率分别为19.4%、10.0%、10.4%和11.4%,其中Cd的点位超标率最高,为7.0%,可以被确定为中国土壤的首要污染物^[7],其污染修复迫在眉睫。但目前,物理、化学和生物修复技术都存在一定的局限,限制了其应用^[8]。某些大型真菌对土壤中的Cd具有很强的富集和耐受能力,利用大型真菌对土壤Cd的修复具有可能性和实践意义,深入揭示大型真菌对重金属的富集和耐性机制也是研究的热点之一。在长期进化过程中,大型真菌已具有相应的Cd胁迫解毒机制,主要包括胞外配体和细胞壁结合及胞内Cd-螯合物形成、抗氧化系统相应和区室化等作用,但其解毒机制尚不完全清楚。随着生物技术的发展,大型真菌耐Cd污染的机制已经从生理生化水平上升至分子水平。本文重点综述了大型真菌耐Cd机制及转录组学技术在系统研究大型真菌耐Cd机制的应用,以期大型真菌耐Cd机制的系统阐明及应用于土壤Cd污染修复实践提供参考。

1 土壤Cd污染现状与修复技术

目前,世界各国土壤都不同程度受到Cd污染,过去50年中,全球排放到环境中的Cd达到 2.20×10^4 t^[9]。美国的国土污染优先治理项目清单显示,在1 200个土壤调查样点中,有63%的样点受到重金属污染,其中受Cd污染的样点占8%^[10];土耳其^[11]、波兰^[12]等国也存在类似问题。近年来,随着工业化和城市化的快速发展,我国面临的土壤Cd污染问题日益严峻。广西水稻主产区的157个稻田耕层(0-20 cm)土样中,35.03%的Cd含量超过国家土壤环境质量的Ⅱ级标准(GB 15618-1995)^[13]。Liu等评估了中国水稻土中Cd的空间格局,结果表明: Cd的浓度范围为0.01-5.50 mg/kg,其中Cd浓度最高的省份是湖南(0.73 mg/kg)、广西(0.70 mg/kg)和四川(0.46 mg/kg);行政区域中, Cd超标率为33.2%,重污染率为8.6%^[14]。《四川省土壤污染状况调查公报》显示,全省土壤总的点位超标率为28.7%,Cd是四川省土壤污染的主要特征污染物,其点位超标率更是高达20.8%^[15]。2022年3月,根据《中共中央 国务院关于深入打好污染防治攻坚战的意见》,生态环境部制定并发布《关于进一步加强重金属污染防治的意见》中对铅(Pb)、汞(Hg)、Cd、铬(Cr)和砷(As)5种重点重金属污染物排放量实施总量控制^[16]。可见,我国土壤重金属污染的治理尤其是土壤Cd污染的治理已成为迫切需要解决的问题,但高效修复技术一直是研究的热点与难点。

物理、化学和生物修复技术是传统以及现阶段比较常用的土壤Cd污染修复方法,物理化学方法主要包括客土法^[17]、电动修复^[18-19]、土壤冲洗/清洗^[20-21]、化学固化/稳定化^[22-23]等;但物理化学修复技术存在耗时耗力、不适用大面积污染土壤、容易造成土壤性质恶化和养分流失、修复产生的废弃物需要进一步处理及存在二次污染的风险等问题^[24-25]。生物修复被认为是一种具有较高发展前景的修复方法,目前研究主要集中在微生物修复^[26]、植物修复^[27-28]及两者的联用^[29]上;但植物修复技术存在有限的修复能力和处理深度、修复周期长、植物和土壤需要长期监测等缺点^[8,22,25];普通的微生物一般肉眼难以看清,对修复环境要求较高,通常难以将重金属从土壤

中彻底去除,两者联合修复依然存在两者独立修复时的缺点。我国土壤类型多样,土壤Cd污染场地特征变异性高,目前的Cd污染土壤修复技术在应用于修复实践过程中均存在一定的局限性。因此,我们需要寻求更为高效、经济和无害的修复方式,大型真菌对Cd的高积累作用使得其成为研究的热点之一。

2 大型真菌对Cd的富集作用

大型真菌是指能形成肉眼可见的子实体、子座、菌核或菌体的一类真菌^[30],包括地衣型真菌、大型子囊菌和担子菌等。大型真菌多样性丰富,世界范围内大型真菌(子囊菌和担子菌)有21 679种^[31];据《中国生物多样性红色名录——大型真菌卷》评估报告,截至2018年,已报道大型子囊菌、担子菌和地衣型真菌共9 302种^[30]。早在20世纪70-80年代,蘑菇属真菌体内的高Cd积累的现象使得大型真菌对重金属超积累作用被发现^[32]。许多大型真菌子实体能有效地吸收和积累土壤中的重金属^[33];且比绿色植物更容易积累高浓度的Cd、Pb和Hg等重金属^[34]。如大型真菌虎皮香菇(*Lentinus tigrinus* (Bull.) Fr.)、鹅膏菌(*Amanita solitaria*)、松乳菇(*Lactarius deliciosus* (L.) Gray)和大白菇(*Russula delica* Fr.)的生物积累系数分别为13.8、30.0、21.8和17.1,高于采摘自同一地点的番茄、胡椒和玉米,其生物积累系数分别是0.048、0.072、0.116^[35]。表1列举了部分对Cd具有较高积累量的大型真菌。

大型真菌积累Cd的能力受种的特异性影响,某些大型真菌会比其他大型真菌具有更强的Cd积累能力。目前已知的野蘑菇、白林地蘑菇、美味牛肝菌和四孢蘑菇对Cd积累量较高,被认为是Cd的超积累真菌^[28]。意大利学者Cocchi分析了食用大型真菌60个种共1 194个样品的Cd含量,最高的是蘑菇属的4个种,鲜重为27.70-101 mg Cd/kg,其中大孢蘑菇101 mg Cd/kg FW和白林地蘑菇39.6 mg Cd/kg FW;皱盖罗鳞伞和菱红菇中Cd的含量分别为24.2和11.6 mg Cd/kg FW,也在Cd高含量之列^[38]。西班牙卢戈省24种野生食用菌中Cd的BCF均大于1,但大孢蘑菇和白林地蘑菇子实体Cd含量最高,分别为 52.9 ± 20.1 和 12.2 ± 3.41 mg/kg DW, BCF分别高达1 399和701,含量最低的是囊状秃马勃[*Calvatia utriformis* (Bull.) Jaap],仅 0.63 ± 0.27 mg/kg, BCF为10.1^[37]。国内凉山彝族自治州的13种野生大型真菌中小白蚁伞和灰褐牛肝菌子实体中Cd含量最高,分别为66.50和11.70 mg/kg DW,而大白菇(*R. delica* Fr.)和黄毛乳菇(*Lactarius representaneus* Britz.)含量较低,仅分别为1.63和1.87 mg/kg DW^[36]。巴西蘑菇对Cd也具有较强的富集作用,在培养料中Cd含量仅为0.17 mg/kg时,其子实体中的Cd含量最高可达5.51 mg/kg, BCF高达32.4^[44]。灰褐牛肝菌中Cd含量为8.69-24.19 mg/kg DW,对应的土壤基质中总Cd含量为0.03-0.57 mg/kg DW,其BCF为24.1-386.0^[42]。大型真菌中富集Cd的含量也与其生长位置是否污染有关。野蘑菇中Cd含量在未受污染的位点为1.0-20 mg/kg DW,污染位点为2.0-50 mg/kg DW^[38,46-49]。美味牛肝菌中Cd含量在未污染和污染位点分别为1.0-10.0和1.0- > 50 mg/kg,极值可高达126.0 mg/kg DW^[38,46-56]。真菌积累重金属部位也与真菌的形态学部位有关。波兰森林中毒蝇鹅膏菌菌盖和菌柄中Cd含量分别为 22.0 ± 11.0 和 8.8 ± 3.9 mg/kg DW, BCF中值:菌盖(1 000) > 菌柄(500)^[40]。人工栽培的大型真菌对Cd的积累量与生长基质中的Cd含量有关,在培养料无污染的正常栽培下的食用菌并不会超过国家规定的标准。在香菇的培养基中分别添加0.07 (CK)、1.14、2.87、5.70和7.87 mg/kg

表1 具有较高Cd积累量的大型真菌

Table 1 Macrofungi with higher Cd accumulation

拉丁名 Latin name	中文名 Chinese name	子实体内Cd含量 Cd content in fruiting body (w/mg kg ⁻¹ , DW)	富集系数 Bioconcentration coefficient (BCF)	生长背景 Growth background	参考文献 References
<i>Amanita pantherina</i>	豹纹鹅膏	11.70	-	野生型 Wild	[36]
<i>Boletus griseus</i>	灰褐牛肝菌	40.60	-	野生型 Wild	[36]
<i>Russula delica</i>	美味红菇	47.20	-	野生型 Wild	[36]
<i>Termitomyces microcarpus</i>	小白蚁伞	135.0	-	野生型 Wild	[36]
<i>Agaricus campestris</i> L.	四孢蘑菇	0.97 ± 0.58	12.30	野生型 Wild	[37]
<i>Amanita rubescens</i> Pers.	赭盖鹅膏菌	0.90 ± 0.43	27.10	野生型 Wild	[37]
<i>Tricholoma equestre</i> (L.) P. Kumm.	油口蘑	0.61 ± 0.30	19.00	野生型 Wild	[37]
<i>Agaricus macrosporus</i> (F.H. Møller & Jul. Schaff.) Pilat	大孢蘑菇	52.90 ± 20.10	1399.00	野生型 Wild	[37]
<i>Agaricus sylvicola</i> (Vittad.) Peck	白林地蘑菇	12.2 ± 3.41	701.00	野生型 Wild	[37]
<i>Rozites caperatus</i> (Pers.: Fr.) P. Karsten	皱盖罗鳞伞	24.20 (鲜重, Fresh weight, FW)	-	野生型 Wild	[38]
<i>Russula vesca</i> Fr.	菱红菇	11.60 (FW)	-	野生型 Wild	[38]
<i>Agaricus arvensis</i> Schff.: Fr.	野蘑菇	325.00	-	野生型 Wild	[39]
<i>Cortinarius caperatus</i> (Pers.) Fr.	皱盖丝膜菌	58.60	-	野生型 Wild	[39]
<i>Laccaria amethystina</i> (Huds.) Cooke	紫蜡蘑	104.00	-	野生型 Wild	[39]
<i>Macrolepiota procera</i> (Scop.) Singer	高大环柄菇	47.30	-	野生型 Wild	[39]
<i>Boletus edulis</i> Bull.	美味牛肝菌	28.40	-	野生型 Wild	[39]
<i>Amanita muscaria</i> (L.) Lam.	毒蝇鹅膏菌	27.10/22.00 ± 11.00	-	野生型 Wild	[39-40]
<i>Thelephora penicillata</i>	帚革菌	1.17 ± 0.37	-	野生型 Wild	[41]
<i>Boletus griseus</i> Forst.	灰褐牛肝菌	8.69-24.19	24.10-386.00	野生型 Wild	[42]
<i>Agaricus blazei</i> Murrill	巴西蘑菇/姬松茸	5.51	32.40	栽培 Cultivated	[43]
<i>Stropharia rugosoannulata</i>	大球盖菇	2.15 ± 0.68	4.29	栽培 Cultivated	[44]
<i>Lentinus edodes</i>	香菇	1.64-75.92	7.04-13.67	栽培 Cultivated	[45]

(实测)的Cd, 武香菌株子实体中Cd含量分别为 1.25 ± 0.27 、 15.34 ± 0.89 、 44.62 ± 1.91 、 44.62 ± 1.91 和 152.78 ± 7.84 mg/kg, BCF分别为18.38、13.46、15.55、16.43和19.41, 且Cd含量分布为菌褶>菌盖>菌柄^[57]。在0.5、5、10 mg/kg Cd胁迫下大球盖菇子实体中Cd含量, 分别为 2.15 ± 0.68 、 5.66 ± 1.13 和 8.10 ± 0.51 mg/kg, BCF分别为4.29、1.13和0.91^[43]。此外, 蘑菇被认为比植物具有更强的环境适应能力和更短的生命周期^[58]; 在废物-作物系统中种植蘑菇修复食物链中Cd也是一种经济高效的方法^[59]。利用大型真菌进行生物修复具有一定的实践意义, 深入揭示大型真菌对重金属的耐受机制是目前重金属污染土壤真菌修复研究的热点之一。

3 大型真菌耐受Cd污染机制

大型真菌同植物一样, 对Cd污染具有一定的适应性, 是多方面作用的结果, 主要可分为胞内作用和胞外作用(图1), 胞外作用主要包括胞外配体固定和细胞壁结合, 减少胞内对Cd的吸收; 胞外的Cd经细胞膜上金属转运蛋白转运进入胞内后, 多种胞内作用主要包括抗氧化系统的防护、Cd-螯合物形成和液泡的区室化, 可降低或消除其毒性^[60-62], 部分Cd还可以经过金属转运蛋白外排出细胞。

3.1 胞外组合作用

真菌对重金属的胞外作用是指胞外组分对重金属离子的络合或螯合^[8], 主要功能物质有有机酸、胞外聚合物(EPS)等^[62]。EPS是原核生物和真核生物产生的^[63], 由脂类、核酸、蛋白质和碳水化合物组成的, 含有羟基(—OH)、羧基(—COOH)和巯基(—SH)等官能团, 使其具有丰富的金属结合位点^[67-68], EPS与重金属离子的作用机制包括电位改变、离子交换、微沉淀和质子交换等^[69]。在10和25 mg/L Cd²⁺胁迫下, 黄孢原毛平革菌EPS去除Cd量占菌体去除Cd量的比例最高, 分别为37.4%和39.3%; 经提取EPS后, 其菌体表面成

分中Cd含量从5.14%下降至4.42%, 进一步表明EPS在黄孢原毛平革菌去除Cd过程中发挥着一定的作用^[70]。扫描电子显微镜(SEM)分析表明暴露于2.4和5.0 mg/L CdCl₂的羊肚菌(*Morchella spongiosa*) M12-10菌丝细表面和细胞壁上有不规则的沉积物, EDS光谱证实这些沉淀物是Cd与细胞结合^[71]。暴露于0.1 mmol/L Cd(NO₃)₂的黄孢原毛平革菌菌丝表面发现类似的Cd晶体颗粒^[72]。重金属胁迫可导致真菌或植物分泌大量低分子量有机酸、EPS等物质, 通过胞外螯合和沉淀等作用, 影响重金属的生物有效性^[8, 73]。Li等研究发现生长黑皮鸡枞菌(*Oudemansiella radicata*)的土壤根际柠檬酸、草酸和苹果酸的含量随Cd胁迫含量的增加而增加, 表明这些酸的产生可能与土壤中Cd的存在密切相关^[74]。

3.2 细胞壁结合作用

细胞壁是重金属进入真菌细胞的第一屏障, 表面带有负电荷, 可通过共价键、静电引力、离子交换等作用吸附重金属离子^[8]。同时细胞壁上也含有多种可与重金属离子结合的活性基团, 能在细胞表面形成络合物或螯合物^[75], 从而阻止过多的重金属进入体内。姬松茸中Cd主要富集在细胞壁上, 其占比高达83.2%, 仅有少部分进入真菌内部^[76]。Cd胁迫下, 水生丝孢菌(*Heliscus lugdunensis*)的菌丝外存在沉淀; 经X射线衍射(XRD)分析确定为非晶颗粒, 能量色散X射线能谱仪分析确定这些非晶颗粒主要由原子比为1:1的Cd和S组成^[77]。细胞壁结构多糖在粗柄侧耳(*Pleurotus platypus*)对Cd的生物吸附中起主导作用; 透射电镜证实金属主要沉积在细胞壁中, 傅里叶变换红外光谱分析证实了—OH、—NH和羰基(—C—O—C)基团参与了Cd的吸附; 金属吸收前后的能量色散X射线分析证明主要吸附机制是离子交换; CaCl₂对Cd解吸的有效性表明Ca²⁺与Cd的交换^[78]。

3.3 Cd结合蛋白结合作用

大型真菌细胞存在一些蛋白或多肽类, 如类金属硫蛋白肽^[79]、金属硫蛋白(MTs)^[80]和谷胱甘肽(GSH)^[81]等, 可与金

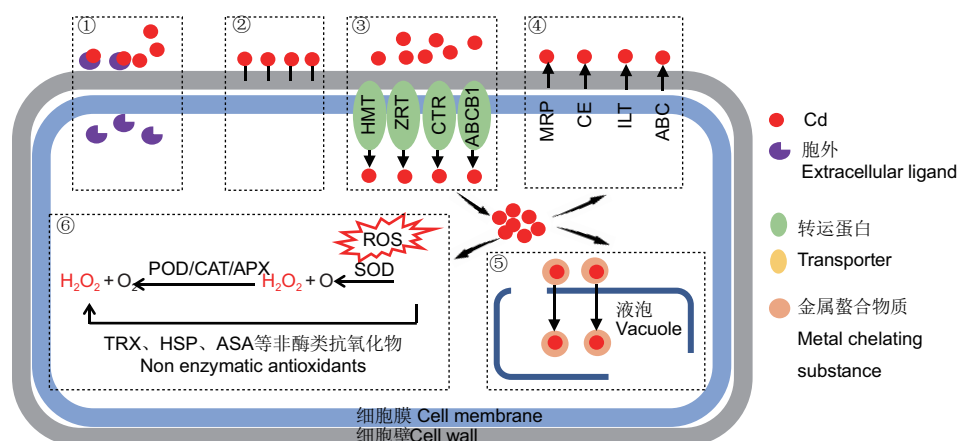


图1 大型真菌对Cd的耐受机制(根据参考文献[61, 63-66]修改绘制)。①胞外配体固定:主要是胞外聚合物(EPS)和低分子量有机酸等可以结合土壤中游离Cd减少其吸收;②细胞壁结合:其含有的多种活性基团如羟基(—OH)、亚氨基(—NH)和羰基(—C—O—C)等可与Cd结合;③Cd通过细胞膜上转运蛋白等进入原生质体;④外排作用:进入细胞内的Cd部分可通过金属转运蛋白从细胞中排出;⑤细胞内整合及液泡“区室化作用”:Cd进入细胞后可能与金属硫蛋白(MTs)、谷胱甘肽(GSH)等整合剂结合,并转运至液泡封存;⑥抗氧化系统(酶类和非酶类系统)的防护作用:负责清除Cd毒害诱导产生的大量ROS,维持细胞内稳态。HMT:重金属转运蛋白;ZRT:锌转运蛋白;CTR:铜转运蛋白;ABCB1:ATP结合盒(ABC)转运体;MRP:多药耐药蛋白;CE:离子转运ATP酶;ILT:铁转运多铜;ABC:ATP结合盒转运蛋白;ROS:活性氧;POD:过氧化物酶;CAT:过氧化氢酶;APX:抗坏血酸过氧化物酶;SOD:超氧化物歧化酶;TRX:硫氧还蛋白;HSP:热休克蛋白;ASA:抗坏血酸。

Fig. 1 Tolerance mechanism of macrofungi to Cd (modified according to references[61] and[63-66]). ① Extracellular ligand fixation: Mainly EPS and low molecular weight organic acids, can combine with free Cd in soil to reduce its absorption; ② Cell wall binding: It contains a variety of active groups (—OH, —NH and —C—O—C) that can bind with Cd; ③ Cd enters the protoplast through membrane transporters (HMT, ZRT, CTR, ABCB1, etc.); ④ Efflux: The part of Cd entering the cell can be discharged from the cell through metal transporters (MRP, CE, ILT and ABC); ⑤ Intracellular chelation and compartmentalization of vacuolar: After Cd enters the cell, it may combine with MTs, GSH and other chelators and transport to the vacuole for storage; ⑥ Protective role of antioxidant system (enzyme and non enzyme system): Responsible for clearing a large amount of ROS induced by Cd toxicity and maintaining intracellular homeostasis. HMT: Heavy metal transporter; ZRT: Zinc transporter; CTR: Copper transporter; ABCB1: ABC-type xenobiotic transporter; MRP: Multidrug resistance protein; CE: Cation transport ATPase; ILT: Iron transport multicopper; ABC: ATP binding cassette transporter; ROS: Active oxygen; POD: Peroxidase; CAT: Catalase; APX: Ascorbate peroxidase; SOD: Superoxide dismutase; TRX: Thioredoxin; HSP: Heat shock protein; ASA: Ascorbic acid.

属离子结合形成硫肽复合物,或封存于液泡等细胞器中,或运输到胞外,以降低重金属的毒性^[8, 82-85]。MTs是一类富含半胱氨酸蛋白质,具有结合金属离子的能力^[86-87]。金属硫酮中存在两个结构域:含有11个半胱氨酸残基的 α -结构域有可能与4个 Cd^{2+} 反应,而包括9个半胱氨酸的 β -结构域可能整合3个 Cd^{2+} ^[88]。GSH是由谷氨酸、半胱氨酸和甘氨酸结合而成的三肽^[89]。10 mg/kg土壤Cd胁迫茶树菇(*Agrocybe aegerita*)菌盖和菌柄的CSH浓度分别为比对照组提高129%和117%;20 mg/kg土壤Cd胁迫下菌盖和菌柄中T-SH的最大浓度分别是对照组的1.51倍和1.38倍,表明Cd与硫醇化合物的络合是Cd细胞解毒的一个途径^[90]。刺囊皮伞(*Cystoderma carcharias*)子实体细胞内Cd的解毒也可能不是源于植物螯合肽(PCs),而是主要依赖于MTs或类MT-肽^[80]。香菇含有一种不含半胱氨酸残基的Cd结合蛋白(LECBP)^[91];通过构建pET28a-Lecbp表达载体及大肠杆菌工程菌,确定了重组LECBP成功表达;IPTG诱导处理后,工程菌对Cd的积累能力显著增强,为对照组的5.2-21.0倍^[92]。

3.4 区室化作用

大型真菌可将进入细胞内重金属离子通过“区室化作用”分布在不同亚细胞部位,从而将其封存或转变成为低毒的形式。茶树菇中53%-75% Cd存在于可溶部分,19%-42% Cd与细胞壁结合,且Cd在细胞壁中的比例随着Cd胁迫浓度的增加而增加^[90]。低浓度(0-0.15 mg/L)和高浓度Cd(2.4-5.0 mg/L)对羊肚菌菌丝细胞影响的最大差异是形成大液泡和细胞质收缩,高Cd处理下整个细胞质区域出现明显的液泡化,其中还有暗电子致密沉积物,这表明液泡可能在限制高浓度Cd的毒性方面发挥重要作用^[71]。分离于污染位点中生黏滑菇

(*Hebeloma mesophaeum*)主要将Cd储存于液泡中,使其具有更高耐Cd性,其可能作用机制与细胞溶质中高含量MTs复合物有关;同时污染位点分离的中生黏滑菇菌株中HmMT3的基础转录水平较清洁位点分离的菌株高了3倍^[93]。Nagy等发现酵母可通过液泡膜上的转运蛋白Yor1p将与PCs或GSH结合的Cd转运到液泡内储存降低其毒性^[94]。

3.5 抗氧化作用

Cd离子进入细胞内,会诱导产生超过细胞中和能力的过量的活性氧(ROS),从而引起氧化应激和诱导病理过程,并最终致使细胞死亡^[8, 95-96]。由抗氧化物质和抗氧化酶组成的细胞抗氧化系统,能够清除ROS自由基,减轻其损害,在重金属耐性中有重要作用^[61, 96]。过氧化氢酶(CAT)、超氧化物歧化酶(SOD)、过氧化物酶(POD)、过氧化物酶(APX)属于重金属胁迫诱导的抗氧化酶^[95, 97-98]。SOD是抵抗ROS的第一道防线,可将超氧阴离子(O_2^-)分解为 H_2O_2 和 O_2 ;CAT和POD将 H_2O_2 分解为 H_2O 和 O_2 ^[99];APX是抗坏血酸(ASA)-GSH循环清除ROS的重要物质^[100]。通常,重金属浓度对抗氧化酶的活性存在“低促高抑”的现象^[101]。此外,也具有一些非酶类的抗氧化物质,如GSH、游离脯氨酸、抗坏血酸(ASA)、丙二醛(MDA)等^[102]。MDA含量是膜脂质过氧化的最终分解产物,是细胞内脂质过氧化的重要指标^[101],脂质膜过氧化可引起细胞内细胞膜损伤^[103]和渗透压平衡失调,而游离脯氨酸可以维持渗透平衡,防止电解质泄漏^[8],并减少ROS的含量^[104]。10 mg/kg土壤Cd胁迫下茶树菇菌盖和菌柄中MDA的含量较对照分别高41%和36%,20 mg/kg土壤Cd胁迫下较对照分别高1.52和1.92倍,说明Cd胁迫下茶树菇细胞膜脂质被氧化;SOD、CAT和POD活性也高于对照,表明这些酶参与茶树菇

胞内Cd的解毒^[90]。低浓度的Cd胁迫不会导致平菇(*Pleurotus ostreatus*) HAU-2菌丝中H₂O₂和MDA显著变化,但是高浓度的Cd胁迫使得菌丝中H₂O₂和MDA显著增加;菌丝中SOD、CAT和POD浓度随Cd胁迫浓度先升高后降低,表明氧化酶可以被相对低浓度的Cd诱导,从而起到去除活性氧的作用;浓度过高的Cd可能严重损伤细胞,从而抑制酶的合成^[105]。类似的结果也出现在Cd胁迫生下的杏鲍菇(*Pleurotus eryngii*)菌丝中,此外,杏鲍菇菌丝中的谷胱甘肽过氧化物酶(GPX)和APX活性表现出类似的趋势^[106]。

4 转录组学技术在大型真菌耐Cd机制研究中的应用

真菌本身是一个复杂的系统,其耐Cd机制是涉及复杂的生理生化过程,传统的研究手段难以对生物体响应胁迫的时效性和动态性研究^[8]。转录组是指在某一功能状态下特定细胞、组织转录出来的所有RNA的总和,包括mRNA与非编码RNA;转录组学是指从整体水平上研究某一阶段特定细胞、组织中全部转录本的转录情况及其转录调控规律的科学^[107]。通过转录组分析可以揭示基因表达和生命活动之间的内在联系^[108, 109]。目前,关于重金属胁迫响应下大型真菌的转录组学研究开展较少。有研究表明:平菇子实体在不同Cd水平处理下的基因表达差异明显;其积累Cd的机制可能与细胞壁吸附、细胞内结合和液泡储存有关;主要涉及通路有细胞壁重塑、金属转运、活性氧反应和富含半胱氨酸相关基因结合等^[110]。利用比较转录组,评估了平菇和黄白侧耳(*Pleurotus cornucopia*)对Cd污染的反应,结果表明:在1和20 mg/L Cd处理下,黄白侧耳中鉴定出705个差异表达基因(DEGs),而在平菇中鉴定出12 551个DEGs;ATP结合盒转运体(ABCB1和线粒体转运体ATM)参与了黄白侧耳的Cd转运,平菇中内吞和吞噬途径得到了增强;26种酶(包括CAT、类过氧化物酶和SOD)在平菇中上调,而在黄白侧耳中仅胞质CAT表达上调,表明两者存在不同的Cd解毒途径。此外,有丝分裂原激活的蛋白激酶信号通路参与了这两个物种的Cd抗性,而不是GSH代谢^[65]。对0.15、0.90和1.50 mg/L Cd²⁺胁迫下羊肚菌菌丝转录组测序,结果表明:共发现了7 444个DEGs,DEGs的数量呈浓度依赖性增加;基于对DEGs的鉴定分析,提出了一个初步的基因调控网络来阐明羊肚菌中Cd解毒的分子机制:羊肚菌菌丝以多种方式增加其Cd耐受性,Cd主要通过重金属转运蛋白(HMT)、锌转运蛋白(ZRT)和铜转运蛋白(CTR)金属转运体的离子通道进入细胞,这导致编码抗氧化酶(SOD、POD、CAT和APX)和非酶抗氧化剂硫氧还蛋白(TRX)和热休克蛋白(HSP)的基因表达增加;参与GSH代谢的基因也被上调,以进一步增强细胞内的活性氧清除;此外,编码金属螯合/结合蛋白(如MT和HMT)的基因上调以隔离Cd并降低其毒性;同时,细胞内转录因子(包括myb和btf)的表达上调,可能影响相关信号转导途径(包括MAPK和cGMP途径)的活性;最后,通过上调膜结合金属转运体[多药耐药蛋白(MRP)、离子转运ATP酶(CE)、铁转运多铜(ILT)和ATP结合盒转运蛋白(ABC)]的表达,Cd从细胞中排出;这些机制的综合作用使羊肚菌菌丝对Cd胁迫具有高度抗性^[66]。0.1 mg/L Cd暴露7 h后,基因型Le4625(具有较高的富集Cd能力)的香菇菌丝中Cd含量为1.390 ± 0.098 mg/kg,约为Le4606(具有较低的富集Cd的能力)的3倍(0.440 ± 0.038

mg/kg);通过RNA Seq评估了总共24 592个转录本,并分析了DEGs,结果表明与Ld4625相比,Ld4606显示出更多的Cd诱导的转录变化;在Ld4606中,Cd暴露后的DEGs与跨膜转运、GSH转移和细胞色素P450相关,表明这些基因可能参与了香菇的Cd抗性。Le4606和Le4625的DEGs编码参与多种生物途径的蛋白质,包括膜上的转运蛋白、细胞壁修饰、氧化应激反应、翻译、降解和信号通路;并且细胞壁修饰、跨膜转运蛋白和金属结合基因在两种基因型中表现出不同的表达模式;细胞中Cd的富集可能激活MAPK信号和抗氧化应激反应,从而改变信号转导和细胞内氧化/还原平衡^[111]。对0、2和5 mg/L Cd胁迫下的巴西蘑菇低Cd积累和耐Cd(J77)菌株和高Cd积累和Cd敏感(J1)菌株的比较转录组分析,结果表明J77菌丝体中鉴定出的Cd响应基因(DEGs)远少于J1菌丝体中的基因(如ABC转运蛋白、ZIP-Zn转运蛋白、GSH S-转移酶和阳离子流出(CE)家族);J1菌丝体中较高的Cd积累可能是由于Cd诱导的ZIP-Zn转运蛋白的上调;细胞内Cd刺激含硫化合物(Cys、Met、GSH、MT2)、细胞壁多糖、有机酸、海藻糖、ATP和NADPH的产生及Cd的封存可能是J1菌丝对Cd积累增加的适应性反应^[112]。以上结果表明细胞壁修饰、抗氧化物质合成、跨膜转运蛋白金属结合相关基因上调可能是大型真菌对Cd富集和耐受的共同途径,信号转导途径激活也在其中起着重要作用,但其不同种和基因型的大型真菌对Cd的耐受机制也存在差异,还有待进一步研究。

5 结论与展望

众多研究已证明多种大型真菌对Cd乃至其他重金属具有极高的富集能力和耐受能力,表明大型真菌作为修复材料在土壤Cd污染修复方面具有应用价值。但大型真菌对于Cd的耐性机制存在种和基因型的特异性,单一的机制并不能解释大型真菌对Cd的耐受性。目前,国内外关于各种大型真菌耐Cd机制的研究也逐渐增多,但大型真菌对Cd的吸附、吸收、转运及耐受机制等方面的研究涉及多个学科,如真菌学、生态毒理学、土壤学、生物化学、分子生物学和生物信息学等,同时也涵盖基因调控、转录控制、代谢调节和生理响应等多个层次,目前的研究较为分散且不成系统。要想将真菌修复土壤Cd污染技术应用于实践,需要多学科交叉进行系统且深入的研究,可在以下几个方面加大研究力度:

(1) 目前对大型真菌对Cd富集过程及耐性机制尚未研究清楚。未来可将生化分析与多组学技术(基因组学、转录组学、蛋白质组学和代谢组学等)联合应用,以提供Cd胁迫下大型真菌基因转录、蛋白质翻译、代谢调控、生理响应及环境影响等方面的详尽信息,全面、系统地阐明大型真菌对Cd胁迫的富集过程和耐受机制存在的共性及特性。

(2) 不同种的大型真菌的种类对Cd的耐受能力存在差异,加强对具有富集Cd能力的大型真菌的筛选、驯化及人工种植研究,丰富修复材料的可选择性;也可向基因工程方面深入研究,如重组和构建某些与重金属转运、解毒相关的基因或者蛋白等,提高大型真菌对Cd的富集和耐受能力,以应对复杂的Cd污染场地特征,增加其应用的普适性。

(3) 探索在各种农业及工业废弃物上种植大型真菌的可能性,以加速真菌修复应用于实践的进程;更远的一步可强化土壤修复后大型真菌子实体的进一步利用及Cd回收等方面的研究,实现资源的循环利用,使真菌修复过程更加经济。

参考文献 [References]

- 1 Akar M, Atis I. The effects of priming pretreatments on germination and seedling growth in perennial ryegrass exposed to heavy metal stress [J]. *Fresenius Environ Bull*, 2018, **27** (10): 6677-6685
- 2 Spiridonova E, Ozolina N, Nesterkina I, Gurina V, Nurminsky V, Donskaya L, Tretyakova A. Effect of cadmium on the roots of beetroot (*Beta vulgaris* L.) [J]. *Int J Phytorem*, 2019, **21** (10): 980-984
- 3 Naeem A, Saifullah U, Rehman MZU, Akhtar T, Ok YS, Rengel Z. Genetic variation in cadmium accumulation and tolerance among wheat cultivars at the seedling stage [J]. *Commun Soil Sci Plant Anal*, 2016, **47** (5): 554-562
- 4 Prozialeck WC, Edwards JR. Mechanisms of cadmium-induced proximal tubule injury: new insights with implications for biomonitoring and therapeutic interventions [J]. *J Pharmacol Exp Ther*, 2012, **343** (1): 2-12
- 5 Person RJ, Tokar EJ, Xu Y, Orihuela R, Ngilame NNO, Waalkes MP. Chronic cadmium exposure *in vitro* induces cancer cell characteristics in human lung cells [J]. *Toxicol Appl Pharmacol*, 2013, **273** (2): 281-288
- 6 Odewumi C, Latinwo L, Sinclair A, Badisa V, Abdullah A, Badisa R. Effect of cadmium on the expression levels of interleukin-1 α and interleukin-10 cytokines in human lung cells [J]. *Mol Med Rep*, 2015, **12** (5): 6422-6426
- 7 环境保护部, 国土资源部. 全国土壤污染状况调查公报[R]. 北京: 环境保护部, 国土资源部, 2014: 1-2 [Ministry of Environmental Protection, Ministry of Land and Resources. Bulletin of National Soil Pollution Survey [R]. Beijing: Ministry of Environmental Protection, Ministry of Land and Resources, 2014: 1-2]
- 8 何海燕. 双孢蘑菇 (*Agaricus bisporus*) 对土壤镉胁迫的抗性机制研究[D]. 北京: 中国科学院大学, 2020 [He HY. Resistance mechanism of *Agaricus bisporus* to cadmium stress in soil [D]. Chinese Academy of Sciences, 2020]
- 9 Singh OV, Labana S, Pandey G, Budhiraja R, Jain RK. Phytoremediation: an overview of metallic ion decontamination from soil [J]. *Appl Microbiol Biotechnol*, 2003, **61** (5): 405-412
- 10 Huang YZ, Hu Y, Liu YX. Combined toxicity of copper and cadmium to six rice genotypes (*Oryza sativa* L.) [J]. *J Environ Sci*, 2009, **21** (5): 647-653
- 11 Coskun M, Steinnes E, Frontasyeva MV, Sjobakk TE, Demkina S. Heavy metal pollution of surface soil in the Thrace region, Turkey [J]. *Environ Monit Assess*, 2006, **119** (1-3): 545-556
- 12 Wiczeorek J, Baran A, Urbański K, Mazurek R, Klimowicz-Pawlas A. Assessment of the pollution and ecological risk of lead and cadmium in soils [J]. *Environ Geochem Health*, 2018, **40** (6): 2325-2342
- 13 陈桂芬, 雷静, 黄雁飞, 熊柳梅, 黄玉溢. 广西稻田镉污染状况及硅对镉的消减作用[J]. 南方农业学报, 2015, **46** (5): 772-776 [Chen GF, Lei J, Huang YF, Xiong LM, Huang YY. Status of heavy metal contamination of paddy soil in Guangxi and effect of silicon fertilizer to reduce Cd content of brown rice [J]. *J S Agric*, 2015, **46** (5): 772-776.
- 14 Liu XJ, Tian GJ, Jiang D, Zhang C, Kong LQ. Cadmium (Cd) distribution and contamination in Chinese paddy soils on national scale [J]. *Environ Sci Pollut R*, 2016, **23** (18): 17941-17952
- 15 四川省人民政府. 四川省土壤污染状况调查公报[EB/OL]. [Sichuan Provincial People's Government Bulletin on Investigation of Soil Pollution in Sichuan Province [E B/OL]] ([2014-12-01] <https://www.sc.gov.cn/10462/10464/10727/10866/2014/12/1/10319883.shtml>)
- 16 中华人民共和国生态环境部. 关于进一步加强重金属污染防治的意见(环固体〔2022〕17号)[EB/OL]. [Ministry of Ecology and Environment of the People's Republic of China. Opinions on further strengthening the prevention and control of heavy metal pollution [EB/OL]] ([2022-03-07] https://www.mee.gov.cn/xgk2018/xgk/xgk03/202203/t20220315_971552.html)
- 17 Douay F, Roussel H, Pruvot C, Lorient A, Fourrier H. Assessment of a remediation technique using the replacement of contaminated soils in kitchen gardens nearby a former lead smelter in Northern France [J]. *Sci Total Environ*, 2008, **401** (1-3): 29-38
- 18 Yeung AT, Gu YY. A review on techniques to enhance electrochemical remediation of contaminated soils [J]. *J Hazard Mater*, 2011, **195**: 11-29
- 19 Giannis A, Gidarakos E. Washing enhanced electrokinetic remediation for removal cadmium from real contaminated soil [J]. *J Hazard Mater*, 2005, **123** (1): 165-175
- 20 Park B, Son Y. Ultrasonic and mechanical soil washing processes for the removal of heavy metals from soils [J]. *Ultrason Sonochem*, 2017, **35**: 640-645
- 21 Wei M, Chen J, Wang X. Removal of arsenic and cadmium with sequential soil washing techniques using Na₂EDTA, oxalic and phosphoric acid: optimization conditions, removal effectiveness and ecological risks [J]. *Chemosphere*, 2016, **156**: 252-261
- 22 Khalid S, Shahid M, Niazi NK, Murtaza B, Bibi I, Dumat C. A comparison of technologies for remediation of heavy metal contaminated soils [J]. *J Geochem Explor*, 2017, **182**: 247-268
- 23 Sun YB, Xu Y, Xu YM, Wang L, Liang XF, Li Y. Reliability and stability of immobilization remediation of Cd polluted soils using sepiolite under pot and field trials [J]. *Environ Pollut*, 2016, **208**: 739-746
- 24 Han R, Dai HP, Yang CJ, Wei SH, Xu L, Yang W, Dou XK. Enhanced phytoremediation of cadmium and/or benzo(a)pyrene contaminated soil by hyperaccumulator *Solanum nigrum* L. [J]. *Int J Phytor*, 2018, **20** (9): 862-868
- 25 Gong YY, Zhao DY, Wang QL. An overview of field-scale studies on remediation of soil contaminated with heavy metals and metalloids: technical progress over the last decade [J]. *Water Res*, 2018, **147**: 440-460
- 26 Huang KW, Dong YP, Chen X, Bie YH, Hou YF, Liao MA. Comparison of cadmium accumulation in three *Solanum* species [J]. *IOP Conference Series: Earth Environ Sci*, 2020, **446** (3): 32008
- 27 Wan XM, Lei M, Yang JX. Two potential multi-metal hyperaccumulators found in four mining sites in Hunan Province, China [J]. *Catena*, 2017, **148**: 67-73
- 28 Wang T, Sun HW, Mao HJ, Zhang YF, Wang CP, Zhang ZY, Wang BL, Sun L. The immobilization of heavy metals in soil by bioaugmentation of a UV-mutant *Bacillus subtilis* 38 assisted by NovoGro biostimulation and changes of soil microbial community [J]. *J Hazard Mater*, 2014, **278**: 483-490
- 29 Ma Y, Oliveira RS, Nai F, Rajkumar M, Luo Y, Rocha I, Freitas H. The hyperaccumulator *Sedum plumbizincicola* harbors metal-resistant endophytic bacteria that improve its phytoextraction capacity in multi-metal contaminated soil [J]. *J Environ Manag*, 2015, **156**: 62-69
- 30 生态环境部, 中国科学院. 《中国生物多样性红色名录——大型真菌

- 卷》评估报告[R]. 北京: 生态环境部, 2018 [Ministry of Ecology and Environment, Chinese Academy of Sciences. Evaluation report of the Red List of Biodiversity in China - Large Macrofungi Volume [R]. Beijing: Ministry of Ecology and Environment, 2018]
- 31 Mueller GM, Schmit JP, Leacock PR, Buyck B, Cifuentes J, Desjardin DE, Halling RE, Hjortstam K, Iturriaga T, Larsson K-H, Lodge DJ, May TW, Minter D, Rajchenberg M, Redhead SA, Ryvarden L, Trappe JM, Watling R, Wu Q. Global diversity and distribution of macrofungi [J]. *Biodivers Conserv*, 2007, **16** (1): 37-48
 - 32 Stijve T, Besson R. Mercury, cadmium, lead and selenium content of mushroom species belonging to the genus *Agaricus* [J]. *Chemosphere*, 1976, **5** (2): 151-158
 - 33 Özcan MM, Dursun N, Al Juhaimi FY. Heavy metals intake by cultured mushrooms growing in model system [J]. *Environ Monit Assess*, 2013, **185** (10): 8393-8397
 - 34 Damodaran D, Vidya Shetty K, Raj Mohan B. Uptake of certain heavy metals from contaminated soil by mushroom-*Galerina vittiformis* [J]. *Ecotoxicol Environ Saf*, 2014, **104**: 414-422
 - 35 Tüzen M. Determination of heavy metals in soil, mushroom and plant samples by atomic absorption spectrometry [J]. *Microchem J*, 2003, **74** (3): 289-297
 - 36 Zhang D, Gao T, Ma P, Luo Y, Su P. Bioaccumulation of heavy metal in wild growing mushrooms from Liangshan Yi nationality autonomous prefecture, China [J]. *Wuhan Univ J Nat Sci*, 2008, **13** (3): 267-272
 - 37 Melgar MJ, Alonso J, Garcia MA. Cadmium in edible mushrooms from NW Spain: bioconcentration factors and consumer health implications [J]. *Food Chem Toxicol*, 2016, **88**: 13-20
 - 38 Cocchi L, Vescovi L, Petrini LE, Petrini O. Heavy metals in edible mushrooms in Italy [J]. *Food Chem*, 2006, **98** (2): 277-284
 - 39 Petkovsek SA, Pokorný B. Lead and cadmium in mushrooms from the vicinity of two large emission sources in Slovenia [J]. *Sci Total Environ*, 2013, **443**: 944-954
 - 40 Falandysz J, Treu R. *Amanita muscaria*: bio-concentration and bio-indicative potential for metallic elements [J]. *Environ Earth Sci*, 2019, **78** (24): 15
 - 41 Borovička J, Braeuer S, Walenta M, Hršelová H, Leonhardt T, Säcký J, Kaňa A, Goessler W. A new mushroom hyperaccumulator: cadmium and arsenic in the ectomycorrhizal basidiomycete *Thelephora penicillata* [J]. *Sci Total Environ*, 2022, **826**: 154227
 - 42 肖俊江, 李鹏程, 庄永亮, 孙丽平, 孙云. 多元统计法分析镉在土壤基质灰褐牛肝菌系统中迁移的影响因素[J]. 分析化学, 2021, **49** (2): 301-308 [Xiao JJ, Li PC, Zhuang YL, Sun LP, Sun Y. Multivariate statistics for investigation of factors affecting migrating of cadmium from soil matrix to *Boletus griseus* [J]. *Chin J Anal Chem*, 2021, **49** (2): 301-308]
 - 43 王丹. 大球盖菇铅、镉富集规律及补硒互作效应研究[D]. 合肥: 安徽农业大学, 2018 [Wang D. Study on the accumulation of heavy metal lead and cadmium and the interaction effect of selenium supplementation in *Stropharia rugosoannulata* [D]. Hefei: Anhui Agricultural University, 2018]
 - 44 黄建成, 应正河, 余应瑞, 李开本. 姬松茸对重金属的富集规律及控制技术研究[J]. 中国农学通报, 2007, **23** (3): 406-409 [Huang JC, Ying ZH, Yu YR, Li KB. Accumulation rule of heavy metal and the controlling technique by *Agaricus blazei* Murrill [J]. *Chin Agric Sci Bull*, 2007, **23** (3): 406-409]
 - 45 徐丽红, 吴应淼, 陈俏彪, 叶长文, 王钢军, 张永志. 香菇 (*Lentinus edodes*) 对重金属镉(Cd)的吸收规律及控制技术研究[J]. 农业环境科学学报, 2011, **30** (7): 1300-1304 [Xu LiH, Wu YM, Chen QB, Ye CW, Wang GJ, Zhang YZ. Investigation of cadmium uptake and accumulation by *Lentinus edodes* and its control technique [J]. *J Agro-Environ Sci*, 2011, **30** (7): 1300-1304]
 - 46 Vetter J. Data on arsenic and cadmium contents of some common mushrooms [J]. *Toxicon*, 1994, **32** (1): 11-15
 - 47 Kalac P, Niznanská M, Bevilacqua D, Stasková I. Concentrations of mercury, copper, cadmium and lead in fruiting bodies of edible mushrooms in the vicinity of a mercury smelter and a copper smelter [J]. *Sci Total Environ*, 1996, **177** (1-3): 251-258
 - 48 Kalac P, Svoboda Lr. A review of trace element concentrations in edible mushrooms [J]. *Food Chem*, 2000, **69** (3): 273-281
 - 49 Kalac P. Trace element contents in European species of wild growing edible mushrooms: a review for the period 2000-2009 [J]. *Food Chem*, 2010, **122** (1): 2-15
 - 50 Kalac P, Burda J, Stasková I. Concentrations of lead, cadmium, mercury and copper in mushrooms in the vicinity of a lead smelter [J]. *Sci Total Environ*, 1991, **105**: 109-119
 - 51 Jorhem L, Sundström B. Levels of some trace elements in edible fungi [J]. *Z Lebensm Unters Forsch*, 1995, **201** (4): 311-316
 - 52 Kraigher H, Al Sayegh Petkovsek S, Grebenc T, Simoncic P. Types of ectomycorrhiza as pollution stress indicators: case studies in Slovenia [J]. *Environ Monit Assess*, 2007, **128** (1-3): 31-45
 - 53 Jamnická G, Bučinová K, Havranová I, Urban A. Current state of mineral nutrition and risk elements in a beech ecosystem situated near the aluminium smelter in Žiar nad Hronom, Central Slovakia [J]. *For Ecol Manag*, 2007, **248** (1): 26-35
 - 54 Frankowska A, Ziółkowska J, Bielawski L, Falandysz J. Profile and bioconcentration of minerals by King Bolete (*Boletus edulis*) from the Płocka Dale in Poland [J]. *Food Addit Contam Part B Surveill*, 2010, **3** (1): 1-6
 - 55 Falandysz J, Frankowska A, Jarzyńska G, Dryżałowska A, Kojta AK, Zhang D. Survey on composition and bioconcentration potential of 12 metallic elements in King Bolete (*Boletus edulis*) mushroom that emerged at 11 spatially distant sites [J]. *J Environ Sci Health, Part B*, 2011, **46** (3): 231-246
 - 56 Giannaccini G, Betti L, Palego L, Mascia G, Schmid L, Lanza M, Mela A, Fabbrini L, Biondi L, Lucacchini A. The trace element content of top-soil and wild edible mushroom samples collected in Tuscany, Italy [J]. *Environ Monit Assess*, 2012, **184** (12): 7579-7595
 - 57 何旭孔, 白冰, 邢增涛, 赵晓燕, 邵毅, 赵明文. 香菇对培养料中镉的富集作用研究 [J]. 食品科学, 2013, **34** (21): 183-187 [He XK, Bai B, Xing ZT, Zhao XY, Shao Y, Zhao MW. Accumulation of cadmium in *Lentinula edodes* from its compost [J]. *Food Sci*, 2013, **34** (21): 183-187]
 - 58 Damodaran D, Balakrishnan RM, Shetty VK. The uptake mechanism of Cd(II), Cr(VI), Cu(II), Pb(II), and Zn(II) by mycelia and fruiting bodies of *Galerina vittiformis* [J]. *BioMed Res Int*, 2013, **2013**: 1-11
 - 59 Stoknes K, Scholwin F, Jasinska A, Wojciechowska E, Mleczek M, Hanc A, Niedzielski P. Cadmium mobility in a circular food-to-waste-to-food system and the use of a cultivated mushroom (*Agaricus subrufescens*) as a remediation agent [J]. *J Environ Manag*, 2019, **245**: 48-54
 - 60 周启星, 安鑫龙, 魏树和. 大型真菌重金属污染生态学研究进展与展望[J]. 应用生态学报, 2008, **19** (8): 1848-1853 [Zhou QX, An XL,

- Wei SH. Heavy metal pollution ecology of macro-fungi: research advances and expectation [J]. *Chin J Appl Ecol*, 2008, **19** (8): 1848-1853
- 61 湛方栋, 何永美, 祖艳群, 李涛. 丝状真菌耐受重金属的细胞机制研究[J]. 云南农业大学学报, 2013, **28** (3): 424-432 [Zhan FD, He YM, Zu YQ, Li T. Cellular mechanisms for heavy metals tolerance of filamentous fungi: a review [J]. *J Yunnan Agric Univ*, 2013, **28** (3): 424-432.
- 62 Yin K, Wang QN, Lü M, Chen LX. Microorganism remediation strategies towards heavy metals [J]. *Chem Eng J*, 2019, **360**: 1553-1563
- 63 Pal A, Paul AK. Microbial extracellular polymeric substances: central elements in heavy metal bioremediation [J]. *Indian J Microbiol*, 2008, **48** (1): 49-64
- 64 刘自然, 甄珍, 陈强, 李玥莹, 王泽, 逢洪波. 植物响应Cd胁迫研究进展[J]. 生物技术通报, 2022, **38** (6): 13-26 [Liu ZR, Zhen Z, Chen Q, Li YY, Wang Z, Pang HB. Research progress in plant response to Cd stress [J]. *Biotechnol Bull*, 2022, **38** (6): 13-26]
- 65 Xu F, Chen P, Li H, Qiao SY, Wang JX, Wang Y, Wang XT, Wu BH, Liu HK, Wang C, Xu H. Comparative transcriptome analysis reveals the differential response to cadmium stress of two *Pleurotus* fungi: *Pleurotus cornucopiae* and *Pleurotus ostreatus* [J]. *J Hazard Mater*, 2021, **416**: 125814
- 66 Xu HY, Xie ZL, Jiang HC, Guo J, Meng Q, Zhao Y, Wang XF. Transcriptome analysis and expression profiling of molecular responses to Cd toxicity in *Morchella spongiosa* [J]. *Mycobiology*, 2021, **49** (4): 421-433
- 67 Wei LL, Li Y, Noguera DR, Zhao NB, Song Y, Ding J, Zhao QL, Cui FY. Adsorption of Cu^{2+} and Zn^{2+} by extracellular polymeric substances (EPS) in different sludges: effect of EPS fractional polarity on binding mechanism [J]. *J Hazard Mater*, 2017, **321**: 473-483
- 68 Shi YH, Huang JH, Zeng GM, Gu YL, Chen YN, Hu Y, Tang B, Zhou JX, Yang Y, Shi LX. Exploiting extracellular polymeric substances (EPS) controlling strategies for performance enhancement of biological wastewater treatments: an overview [J]. *Chemosphere*, 2017, **180**: 396-411
- 69 Guibaud G, Hullebusch EV, Bordas Fo, D'abzac P, Joussein E. Sorption of Cd(II) and Pb(II) by exopolymeric substances (EPS) extracted from activated sludges and pure bacterial strains: modeling of the metal/ligand ratio effect and role of the mineral fraction [J]. *Bioresour Technol*, 2009, **100** (12): 2959-2968
- 70 吴磊, 张学洪, 李宁杰, 陈中维, 兰琪, 刘洁. 胞外聚合物在白腐真菌去除镉过程中的作用[J]. 桂林理工大学学报, 2020, **40** (1): 177-181 [Wu L, Zhang XH, Li NJ, Chen ZW, Lan Q, Liu J. Roles of extracellular polymeric substances in cadmium removal by white rot fungi [J]. *J Guilin Univ Technol*, 2020, **40** (1): 177-181]
- 71 Xu HY, Xie ZL, Guo J, Men Q. Morphological changes and bioaccumulation in response to cadmium exposure in *Morchella spongiosa*, a fungus with potential for detoxification [J]. *Can J Microbiol*, 2021, **67** (11): 789-798
- 72 Chen AW, Zeng GM, Chen GQ, Liu L, Shang C, Hu XJ, Lu LH, Chen M, Zhou Y, Zhang QH. Plasma membrane behavior, oxidative damage, and defense mechanism in *Phanerochaete chrysosporium* under cadmium stress [J]. *Process Biochem*, 2014, **49** (4): 589-598
- 73 何海燕, 张丹, 徐露, 高嘉宁, 向宇国, 王波, 吴毅, 曹龙武. 镉胁迫下双孢蘑菇发酵液中低分子量有机酸分泌量与镉含量变化[J]. 应用与环境生物学报, 2020, **26** (1): 181-189 [He HY, Zhang D, Xu L, Gao JN, Xiang YG, Wang B, Wu Y, Cao LW. Changes in secretion of low molecular weight organic acids and cadmium content in the fermentation broth of *Agaricus bisporus* [J]. *Chin J Appl Environ Biol*, 2020, **26** (1): 181-189]
- 74 Li XD, Xiao KM, Ma H, Li LL, Tan H, Xu H, Li YZ. Mechanisms into the removal and translocation of cadmium by *Oudemansiella radicata* in soil [J]. *Environ Sci Pol Res*, 2019, **26** (7): 6388-6398
- 75 张秀丽, 刘月英. 贵、重金属的生物吸附[J]. 应用与环境生物学报, 2002, **8** (6): 668-671 [Zhang XL, Liu YY. Biosorption of precious and heavy metal [J]. *Chin J Appl Environ Biol*, 2002, **8** (6): 668-671.
- 76 李三暑, 雷锦桂, 颜明娟, 江枝和. 镉对姬松茸细胞悬浮培养的影响及其在细胞中的分布[J]. 江西农业大学学报, 2001, **23** (3): 329-331 [Li SS, Lei JG, Yan MJ, Jiang ZH. The effects of cadmium on suspension culture of *Agaricus blazei* Murrill cells and the distribution of cadmium [J]. *Acta Agric Univ Jiangxiensis*, 2001, **23** (3): 329-331]
- 77 Dobritzsch D, Ganz P, Rother M, Ehrman J, Baumbach R, Miersch J. Cadmium-induced formation of sulphide and cadmium sulphide particles in the aquatic hyphomycete *Heliscus lugdunensis* [J]. *J Trace Elem Med Biol*, 2015, **31**: 92-97
- 78 Vimala R, Das N. Mechanism of Cd(II) adsorption by macrofungus *Pleurotus platypus* [J]. *J Environ Sci*, 2011, **23** (2): 288-293
- 79 Leonhardt T, Sack J, Simek P, Santrucek J, Kotrba P. Metallothionein-like peptides involved in sequestration of Zn in the Zn-accumulating ectomycorrhizal fungus *Russula atropurpurea* [J]. *Metallomics*, 2014, **6** (9): 1693-1701
- 80 Borovicka J, Braeuer S, Sack J, Kameník J, Goessler W, Trubač J, Strnad L, Rohovec J, Leonhardt T, Kotrba P. Speciation analysis of elements accumulated in *Cystoderma carcharias* from clean and smelter-polluted sites [J]. *Sci Total Environ*, 2019, **648**: 1570-1581
- 81 Bellion M, Courbot M, Jacob C, Blaudez D, Chalot M. Extracellular and cellular mechanisms sustaining metal tolerance in ectomycorrhizal fungi [J]. *FEMS Microbiol Lett*, 2006, **254** (2): 173-181
- 82 Benes V, Leonhardt T, Sack J, Kotrba P. Two $\text{P}_{1\text{B-1}}$ -ATPases of *Amanita strobiliformis* with distinct properties in Cu/Ag transport [J]. *Front Microbiol*, 2018, **9**: 747
- 83 Sack J, Leonhardt T, Kotrba P. Functional analysis of two genes coding for distinct cation diffusion facilitators of the ectomycorrhizal Zn-accumulating fungus *Russula atropurpurea* [J]. *Biomaterials*, 2016, **29** (2): 349-363
- 84 Ruytinx J, Nguyen H, Van Hees M, Op De Beeck M, Vangronsveld J, Carleer R, Colpaert JV, Adriaensen K. Zinc export results in adaptive zinc tolerance in the ectomycorrhizal basidiomycete *Suillus bovinus* [J]. *Metallomics*, 2013, **5** (9): 1225-1233
- 85 Blaudez D, Chalot M. Characterization of the ER-located zinc transporter ZnT1 and identification of a vesicular zinc storage compartment in *Hebeloma cylindrosporum* [J]. *Fungal Genetics Biol*, 2011, **48** (5): 0-503
- 86 Freisinger E, Vasak M. Cadmium in metallothioneins [C]//Sigel A, Sigel H, Sigel R K O. Cadmium: from toxicity to essentiality. Dordrecht: Springer Netherlands, 2013: 339-371
- 87 Klaassen CD, Liu J, Diwan BA. Metallothionein protection of cadmium toxicity [J]. *Toxicol Appl Pharm*, 2009, **238** (3): 215-220
- 88 Sutherland DEK, Stillman MJ. Challenging conventional wisdom: single domain metallothioneins [J]. *Metallomics: Int Biol Sci*, 2014, **6** (4): 702-728

- 89 Thévenod F, Lee WK. Toxicology of cadmium and its damage to mammalian organs [C]//Sigel A, Sigel H, Sigel RKO. Cadmium: from toxicity to essentiality. Dordrecht: Springer Netherlands, 2013: 415-490
- 90 Li XD, Ma H, Li LL, Gao YF, Li YZ, Xu H. Subcellular distribution, chemical forms and physiological responses involved in cadmium tolerance and detoxification in *Agrocybe aegerita* [J]. *Ecotoxicol Environ Saf*, 2019, **171**: 66-74
- 91 Dong XB, Liu Y, Feng X, Shi D, Bian YB, Ibrahim SA, Huang W. Purification and characterization of a cadmium-binding protein from *Lentinula edodes* [J]. *J Agric Food Chem*, 2019, **67** (4): 1261-1268
- 92 董晓博. 镉结合蛋白(LECBP)在香菇镉富集中的作用研究[D]. 武汉: 华中农业大学, 2023 [Dong XB. Study on the role of cadmium-binding protein (LECBP) in cadmium enrichment in *Lentinula edodes* [D]. Wuhan: Huazhong Agricultural University, 2023]
- 93 Sackay J, Benes V, Borovicka J, Leonhardt T, Kotrba P. Different cadmium tolerance of two isolates of *Hebeloma mesophaeum* showing different basal expression levels of metallothionein (*HmMT3*) gene [J]. *Fungal Biol-UK*, 2019, **123** (3): 247-254
- 94 Nagy Z, Montigny C, Leverrier P, Yeh S, Goffeau A, Garrigos M, Falson P. Role of the yeast ABC transporter Yor1p in cadmium detoxification [J]. *Biochimie*, 2006, **88** (11): 1665-1671
- 95 Su YQ, Zhao YJ, Wu N, Chen YE, Zhang WJ, Qiao DR, Cao Y. Chromium removal from solution by five photosynthetic bacteria isolates [J]. *Appl Microbiol Biot*, 2018, **102** (4): 1983-1995
- 96 Belozerskaya TA, Gessler NN. Reactive oxygen species and the strategy of antioxidant defense in fungi: a review [J]. *Appl Biochem Microbiol*, 2007, **43** (5): 506-515
- 97 Kumar V, Singh S, Singh G, Dwivedi S. Exploring the cadmium tolerance and removal capability of a filamentous fungus *Fusarium solani* [J]. *Geomicrobiol J*, 2019, **36** (9): 782-791
- 98 Azizollahi Z, Ghaderian SM, Ghotbi-Ravandi AA. Cadmium accumulation and its effects on physiological and biochemical characters of summer savory (*Satureja hortensis* L.) [J]. *Int J Phytorem*, 2019, **21** (12): 1241-1253
- 99 Irfan M, Ahmad A, Hayat S. Effect of cadmium on the growth and antioxidant enzymes in two varieties of *Brassica juncea* [J]. *Saudi J Biol Sci*, 2014, **21** (2): 125-131
- 100 Mishra B, Sangwan RS, Mishra S, Jadaun JS, Sabir F, Sangwan NS. Effect of cadmium stress on inductive enzymatic and nonenzymatic responses of ROS and sugar metabolism in multiple shoot cultures of Ashwagandha (*Withania somnifera* Dunal) [J]. *Protoplasma*, 2014, **251** (5): 1031-1045
- 101 Jiang J, Liu H, Li Q, Gao N, Yao Y, Xu H. Combined remediation of Cd-phenanthrene co-contaminated soil by *Pleurotus cornucopiae* and *Bacillus thuringiensis* FQ1 and the antioxidant responses in *Pleurotus cornucopiae* [J]. *Ecotox Environ Safe*, 2015, **120**: 386-393
- 102 Pandhair V, Sekhon B. Reactive oxygen species and antioxidants in plants: an overview [J]. *J Plant Biochem Biotechnol*, 2006, **15** (2): 71-78
- 103 Yang LP, Zhu J, Wang P, Zeng J, Tan R, Yang YZ, Liu ZM. Effect of Cd on growth, physiological response, Cd subcellular distribution and chemical forms of *Koeleria paniculata* [J]. *Ecotox Environ Safe*, 2018, **160**: 10-18
- 104 Xu J, Yin HX, Li X. Protective effects of proline against cadmium toxicity in micropropagated hyperaccumulator, *Solanum nigrum* L. [J]. *Plant Cell Rep*, 2009, **28** (2): 325-333
- 105 Li XZ, Wang YJ, Pan YS, Yu H, Zhang XL, Shen YP, Jiao S, Wu K, La GX, Yuan Y, Zhang SM. Mechanisms of Cd and Cr removal and tolerance by macrofungus *Pleurotus ostreatus* HAU-2 [J]. *J Hazard Materials*, 2017, **330**: 1-8
- 106 Huang XH, Xu N, Feng LG, Lai DN, Wu F, Xu D, Guo X. The activity and gene expression of enzymes in mycelia of *Pleurotus eryngii* under cadmium stress [J]. *Sustainability*, 2022, **14** (7): 4125
- 107 林海燕, 曾超珍, 谭斌, 王坤波, 刘仲华. 转录组学技术在茶树抗逆性的研究进展[J]. 分子植物育种, 2019, **17** (3): 803-810 [Lin HY, Zeng CZ, Tan B, Wang KB, Liu ZH. Research progress of transcriptome technology in tea plant (*Camellia sinensis*) stress tolerance [J]. *Mol Plant Breeding*, 2019, **17** (3): 803-810]
- 108 宋海凤. 雌雄青杨(*Populus cathayana*)根系对氮素缺乏的差异响应: 基于组学分析[D]. 北京: 中国科学院大学, 2019: 7 [Song HF. Multi-omic approaches in exploring sexually differential responses of *Populus cathayana* roots to nitrogen deficiency [D]. Beijing: Chinese Academy of Sciences, 2019: 7]
- 109 范慧艳. 甜菜坏死黄脉病毒侵染本生烟和大果甜菜的生物学、转录组学和蛋白质组学研究[D]. 北京: 中国农业大学, 2014: 2 [FAN HY. Biological, transcriptomic and proteomic analysis of *Nicotiana benthamiana* and *Beta macrocarpa* infected by *Beet necrotic yellow vein virus* [D]. Beijing: China Agricultural University, 2014: 2]
- 110 Chen MM, Wang LK, Hou JL, Yang SS, Zheng X, Chen L, Li XF. Mycoextraction: rapid cadmium removal by macrofungi-based technology from alkaline soil [J]. *Minerals*, 2018, **8** (12): 589
- 111 Yu H, Li Q, Shen X, Zhang L, Liu J, Tan Q, Li Y, Lv B, Shang X. Transcriptomic analysis of two *Lentinula edodes* genotypes with different cadmium accumulation ability [J]. *Front Microbiol*, 2020, **11**: 558104
- 112 Liu PH, Huang ZX, Luo XH, Chen H, Weng BQ, Wang YX, Chen LS. Comparative transcriptome analysis reveals candidate genes related to cadmium accumulation and tolerance in two almond mushroom (*Agaricus brasiliensis*) strains with contrasting cadmium tolerance [J]. *PLoS ONE*, 2020, **15** (9): e239617